

## RESEARCH

# Long term exposure to ambient air pollution and incidence of acute coronary events: prospective cohort study and meta-analysis in 11 European cohorts from the ESCAPE Project

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## Abstract

**Objectives** To study the effect of long term exposure to airborne pollutants on the incidence of acute coronary events in 11 cohorts participating in the European Study of Cohorts for Air Pollution Effects (ESCAPE).

**Design** Prospective cohort studies and meta-analysis of the results.

**Setting** Cohorts in Finland, Sweden, Denmark, Germany, and Italy.

**Participants** 100 166 people were enrolled from 1997 to 2007 and followed for an average of 11.5 years. Participants were free from previous coronary events at baseline.

**Main outcome measures** Modelled concentrations of particulate matter <2.5 µm (PM<sub>2.5</sub>), 2.5-10 µm (PM<sub>coarse</sub>), and <10 µm (PM<sub>10</sub>) in aerodynamic diameter, soot (PM<sub>2.5</sub> absorbance), nitrogen oxides, and traffic exposure at the home address based on measurements of air pollution conducted in 2008-12. Cohort specific hazard ratios for incidence of acute coronary events (myocardial infarction and unstable angina) per fixed increments of the pollutants with adjustment for sociodemographic and lifestyle risk factors, and pooled random effects meta-analytic hazard ratios.

**Results** 5157 participants experienced incident events. A 5 µg/m<sup>3</sup> increase in estimated annual mean PM<sub>2.5</sub> was associated with a 13% increased risk of coronary events (hazard ratio 1.13, 95% confidence interval 0.98 to 1.30), and a 10 µg/m<sup>3</sup> increase in estimated annual mean PM<sub>10</sub> was associated with a 12% increased risk of coronary events (1.12, 1.01 to 1.25) with no evidence of heterogeneity between cohorts. Positive associations were detected below the current annual European limit value of 25 µg/m<sup>3</sup> for PM<sub>2.5</sub> (1.18, 1.01 to 1.39, for 5 µg/m<sup>3</sup> increase in PM<sub>2.5</sub>) and below 40 µg/m<sup>3</sup> for PM<sub>10</sub> (1.12, 1.00 to 1.27, for 10 µg/m<sup>3</sup> increase in PM<sub>10</sub>). Positive but non-significant associations were found with other pollutants.

**Conclusions** Long term exposure to particulate matter is associated with incidence of coronary events, and this association persists at levels of exposure below the current European limit values.

## Introduction

According to the recent report on the Global Burden of Disease, throughout the world particulate air pollution is estimated to cause 3.1 million deaths a year and 22% of disability adjusted life years (DALY) due to ischaemic heart disease.<sup>1</sup> Several cohort studies have reported that long term exposure to air pollution is associated with mortality, in particular cardiovascular mortality.<sup>2-13</sup> The evidence of an effect on

incidence of cardiovascular events—that is, acute myocardial infarction and unstable angina—is less consistent and requires further investigations.<sup>12-15</sup>

In the European Union the current annual limit for particulate matter <2.5 µm (PM<sub>2.5</sub>) is 25 µg/m<sup>3</sup>, which is far above that implemented in the United States (12 µg/m<sup>3</sup>). One obstacle in the European standard setting process in the past had been that the available estimates of the exposure-response associations of particulate matter were primarily based on studies conducted in North America.<sup>3-12</sup> In response, the ESCAPE Study (European Study of Cohorts for Air Pollution Effects) was conducted between 2008 and 2012 to quantify the associations between exposures and health outcomes by using standardised methods for assessment of exposure and data from existing cohort studies.<sup>16 17</sup>

We estimated the association between long term exposure to particulate matter <2.5 µm (PM<sub>2.5</sub>), 2.5-10 µm (coarse), <10 µm (PM<sub>10</sub>) in aerodynamic diameter, soot (PM<sub>2.5</sub> absorbance), nitrogen oxides (NO<sub>x</sub> and NO<sub>2</sub>), and traffic indicators and the incidence of coronary events. We evaluated effect modification by several individual characteristics, and we investigated the exposure-response relations of the pollutants below selected thresholds.

## Methods

### Design and population

This study is an analysis of cohort data obtained by ESCAPE to investigate the long term effects of exposure to air pollution on human health in Europe and a meta-analysis of the cohort specific results. The present study included 11 European cohorts from five countries with information about incident cases of acute coronary events and the most important potential confounders. The cohorts were in Finland (FINRISK)<sup>18</sup>; Sweden (the Swedish National Study on Aging and Care in Kungsholmen (SNAC-K), the Screening Across the Lifespan Twin Study, (SALT), the 60 year olds study, and the Stockholm Diabetes Prevention Program study (SDPP))<sup>19-22</sup>; Denmark (the Danish Diet, Cancer and Health cohort study [DCH])<sup>23</sup>; Germany (the Heinz Nixdorf Recall Study (HNR), the Cooperative Health Research in the Augsburg Region (KORA))<sup>24 25</sup>; and Italy (the European Prospective Investigation into Cancer and Nutrition

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Extra material supplied by the author (see <http://www.bmj.com/content/348/bmj.f7412?tab=related#webextra>)

**Appendix:** Cohort details and noise assessment

Video on bmj.com (see also <http://bmj.com/video>)



Video abstract

in Turin (Epic-Turin), the International Study on Asthma and Allergies in Childhood in Turin and Rome (parents of representative samples of children participating in SIDRIA-Turin and SIDRIA-Rome)).<sup>26 27</sup> These were all centres included in the ESCAPE intensive monitoring programme for particulate matter and nitrogen oxides.<sup>28 29</sup> Table 1 shows specific settings, periods of recruitment, and person years at risk. With the exception of the Italian cohorts, all the cohorts enrolled representative samples of the adult population for the age groups considered in the single studies. Additional information on each cohort is presented in the appendix.

## Outcome definition

For nine out of 11 cohorts, information on coronary events was derived by record linkage procedures with hospital discharge and mortality registries. We selected international classification of diseases (ICD) codes for “acute myocardial infarction” or “other acute and sub-acute forms of ischemic heart disease” (ICD-9-CM codes: 410, 411; ICD-10 codes: I21, I23, I20.0, I24) in principal diagnoses of hospital discharges. We also considered as incident cases those people who died outside hospital from ischaemic heart diseases, according to the death certificates (ICD-9: 410-414; ICD-10: I20-I25) and had no evidence of hospital admission for ischaemic heart disease in the 28 days before death and no evidence of admission for any cause in the two days before death. To identify incident cases, we excluded those who had had an acute coronary event or cerebrovascular event previously (previous admission to hospital with principal or secondary diagnoses with ICD-9 codes 410, 411, 430, 431, 434, 436 and ICD-10 codes I21, I23, I20.0, I24, I61, I63, I64). The reference period to evaluate previous events varied across the cohorts, with the minimum of three years in the two SIDRIA cohorts and all lifelong history for the KORA cohort. For the HNR study, incident cases were adjudicated by an independent end point committee based on medical records.<sup>24</sup> For the KORA cohort, incident cases were determined by interview and medical history and validated through linkage with the myocardial infarction register.<sup>25</sup>

## Exposure to air pollution

Exposure to air pollution in each area was estimated following a standard methods developed for the ESCAPE project and described elsewhere.<sup>28 29</sup> Briefly, for each area under study, particulate matter of varying sizes measured in  $\mu\text{m}$  (shown as subscript)—that is,  $\text{PM}_{10}$ , coarse PM,  $\text{PM}_{2.5}$ , and  $\text{PM}_{2.5}$  absorbance (blackness of the  $\text{PM}_{2.5}$  exposed filter, determined by measurement of light reflectance as a marker for soot and black carbon)—was measured in 20 sites, and nitrogen oxides were measured in 40 sites in three separate two week periods (to cover different seasons) over one year (between 2008 and 2011). For each site, results from the three measurements were averaged to estimate the annual average, with adjustment for temporal variation by using a centrally located background reference site, which was operated for a whole year.<sup>30 31</sup> By using several traffic and land use variables, we developed area specific land use regression (LUR) models to explain the spatial variation of each measured pollutant. These models were then used to estimate concentrations of air pollution at each participant’s residential address. Geographical variables typically evaluated include altitude, population density, industrial land use, green space, and traffic flows variables.<sup>30 31</sup> In addition to concentrations of pollutants, we considered two traffic variables at the participant’s residence: traffic intensity on the nearest road (vehicles/day) and traffic load on major roads in a 100 m buffer (vehicles $\times$ m/day), defined as the sum of traffic intensity

on roads with >5000 vehicles/day multiplied by the length of those roads in a 100 m buffer. To validate the models, we used the leave one out cross validation method—that is, systematically subtracting each of the monitoring points from the model one by one, and then comparing the predicted value for each monitoring location with the measured level at the location without using this measurement in the development of the model.<sup>30 31</sup>

## Covariates

All cohorts had a common set of potential confounders and effect modifiers at baseline: marital status (coded as single, married/living with partner, divorced/separated, widowed; for the SDPP cohort only the binary variable “living with partner” was available; for SIDRIA-Rome all participants were living as couples at baseline), education (primary school or less, up to secondary school or equivalent, university degree and more), occupation (employed, unemployed, homemaker/housewife, retired), smoking status (current, former, never), duration of smoking (years), smoking intensity (cigarettes/day), hypertension, and diabetes. Most cohorts had information on additional cardiovascular risk factors such as body mass index (BMI; coded as <25, 25-29,  $\geq$ 30), physical activity (<1 hour/week, about 1 hour/week, >2 hours/week), alcohol consumption (never, 1-3 drinks/week, 3-6 drinks/week, >6 drinks/week), and three cohorts had information on use of drugs for hormone replacement therapy. Four of 11 cohorts had data on cholesterol concentrations. Ten of 11 cohorts had data on noise exposure, estimated at the residential addresses from European noise exposure assessment in 2007, and we used an eight class categorical variable of 5 dB of exposure from 45 dB to 75 dB and more (see appendix). Each cohort also used an area based socioeconomic status indicator because socioeconomic indicators at the area level are predictors of morbidity, access to care, and lifestyle risk factors in addition to individual socioeconomic characteristics.<sup>32</sup>

## Statistical analyses

To evaluate the association between exposure to air pollution and incidence of coronary events we performed the analyses in two stages. Firstly, we analysed each cohort using a common protocol for confounders, outcomes, and statistical modelling. We used Cox proportional hazards regression models (hazard ratios) with age as the time scale. A script with statistical code was provided to all cohort specific analysts. The results were evaluated centrally at the Department of Epidemiology in Rome. Secondly, we carried out a random effects meta-analysis to pool results.<sup>33</sup> We calculated  $I^2$  statistics and P values for  $\chi^2$  test from Cochran’s Q to quantify the heterogeneity among studies.<sup>34</sup>

We first estimated hazard ratios adjusted for age, sex, and year of enrolment only (model 1), then adjusted for a common set of individual covariates (marital status, education, occupation, smoking status, smoking duration, and smoking intensity; model 2), and then adjusted for area based socioeconomic status indicators (model 3). When investigating the effects of traffic variables we also adjusted all models for background  $\text{NO}_2$  levels. We estimated the effects of exposure to air pollution using fixed increments of pollutants (20  $\mu\text{g}/\text{m}^3$  for  $\text{NO}_2$ , 10  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{10}$  and  $\text{NO}_2$ , 5  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{2.5}$  and  $\text{PM}_{\text{coarse}}$ ,  $10^{-5}/\text{m}$  for  $\text{PM}_{2.5}$  absorbance).

We performed several sensitivity analyses to deal with potential sources of bias and to consider the confounding role of additional cardiovascular risk factors. Firstly, we adjusted for possible intermediate variables (diabetes, hypertension) available



in all the cohorts; then for physical activity, alcohol consumption, and BMI (available in eight cohorts); finally we added to previously mentioned factors the cholesterol concentration (available in four cohorts). We then analysed the role of living in low/high urbanised settings, of noise exposure, and of residential stability (restricting the analysis to those living at the same baseline address during the follow-up). We stratified the Cox model 3 for predictors that did not meet the proportionality hazard assumption. We analysed the influential role of the largest cohort by excluding the DCH study from the analysis. Finally, we evaluated the role of the performance of land use regression models, stratifying cohorts by cross validation  $R^2$  or  $\leq 60\%$ .

To explore the effect of clustering by area of residence in the association between exposure and mortality—that is, residents in the same area usually share similar characteristics (socioeconomic status, health, access to services) and have similar environmental and air conditions—we performed a frailty model to measure the role each area played.<sup>4</sup>

We evaluated potential effect modification by adding to model 3 an interaction term of exposure and one effect modifier at a time (sex, educational level, smoking status, BMI, hypertension, and residence in low or high urbanised settings). We used the likelihood ratio test to compare the models with and without interaction terms. For age during follow-up, we estimated effect modification by adding an interaction term between exposure and a time dependent categorical variable indicating age group (<60, 60-74, >75).

To provide information about the health effects below specific threshold values (20, 30, and 40  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{10}$ , and 15, 20, and 25  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{2.5}$ ), we studied the effect of fixed increments of  $\text{PM}_{2.5}$  and  $\text{PM}_{10}$  among people with air pollution concentrations at residences only below these thresholds.

We used STATA software (versions 10, 11, and 12) for all the analyses, with the exception of frailty models for which we used R ([www.r-project.org](http://www.r-project.org)).

## Results

Table 1 shows a summary description of the 11 European cohorts. Table 2 shows individual characteristics of participants and table 3 their exposures to air pollution. Additional characteristics of the participants are provided in table 4. The enrolment period covered 15 years, and the average follow-up was 11.5 years. The participating cohorts varied in characteristics, availability of data on covariates, and levels of exposure. The proportion of participants included in the study ranged from 82.4% to 99.3% of the original cohorts (overall 93.1%), after exclusion of missing values on any of the covariates in model 3. There were no differences in exposure levels between included and excluded participants. Among 100 166 participants included in the study and followed for 1 154 386 person years, there were 5157 incident cases. The average  $\text{PM}_{2.5}$  level at residence ranged from 7.3  $\mu\text{g}/\text{m}^3$  in Sweden ( $\text{SD}=1.3$ ) to 31.0  $\mu\text{g}/\text{m}^3$  ( $\text{SD}=1.7$ ) in northern Italy, and all other pollutants had similar patterns. All exposure models had good performance ( $R^2 \geq 0.61$ ), and the prediction ability of models for nitrogen dioxides (based on a larger number of measurements sites) was slightly higher than those of particulate models.

Table 5 shows the pooled hazard ratios with 95% confidence intervals for incidence of coronary events for fixed increments of all pollutants. We observed the strongest association for  $\text{PM}_{10}$ . While in model 1 (adjusted for age, sex, and calendar period) all particulate matter indicators were strongly associated with incidence of coronary events, when we adjusted for marital

status, education, occupation, smoking status, smoking duration, smoking intensity, and socioeconomic area indicator (model 3), only  $\text{PM}_{10}$  showed a significant association (hazard ratio 1.12, 95% confidence interval 1.01 to 1.25, for each 10  $\mu\text{g}/\text{m}^3$  increase). There was also an association for  $\text{PM}_{2.5}$  (1.13, 0.98 to 1.30, for each 5  $\mu\text{g}/\text{m}^3$  increase) and for coarse particles, whereas only small positive associations were found for nitrogen oxides. There was no evidence of an effect of traffic variables. There was no evidence of statistical heterogeneity between the cohort specific effect estimates ( $I^2 < 5\%$ ). The figure shows the forest plots of 10  $\mu\text{g}/\text{m}^3$   $\text{PM}_{10}$  and 5  $\mu\text{g}/\text{m}^3$   $\text{PM}_{2.5}$  increments (model 3).

Table 6 shows the results of the sensitivity analyses for  $\text{PM}_{10}$  and  $\text{PM}_{2.5}$  from the base model (model 3), reporting the number of cohorts and participants included in each analysis. Overall, there were only a few marginal changes in the effect estimates across the various models, and there was no evidence of heterogeneity among cohorts. When we considered diabetes and hypertension in the adjustment, the effect estimates for  $\text{PM}_{10}$  and for  $\text{PM}_{2.5}$  were slightly lower (hazard ratio 1.11 (95% confidence interval 1.00 to 1.24) and 1.11 (0.96 to 1.28), respectively). When we considered physical activity, alcohol consumption, and BMI in the eight cohorts with available information, the association between particulate matter and incidence of coronary events remained stable. Adjustment for cholesterol in the four cohorts with this variable did not alter the results. Adjustment for location of residence (urban, suburban/rural) slightly increased the effect estimates. In 63 121 participants who did not change their address during the follow-up, the effect estimate of air pollution was higher than in the whole population (hazard ratio 1.16 (1.01 to 1.32) for 10  $\mu\text{g}/\text{m}^3$   $\text{PM}_{10}$ , and 1.18 (0.98 to 1.42) for 5  $\mu\text{g}/\text{m}^3$   $\text{PM}_{2.5}$ ), possibly because of more accurate exposure assignment. Exclusion of the large DCH study did not influence the results. When we restricted the meta-analysis to nine cohorts with exposure models with high cross validation ( $R^2 > 60$ ) for  $\text{PM}_{10}$  and six cohorts for  $\text{PM}_{2.5}$ , we found stronger associations (hazard ratio 1.18 (1.05 to 1.33) for 10  $\mu\text{g}/\text{m}^3$   $\text{PM}_{10}$ , and 1.35 (1.04 to 1.74) for 5  $\mu\text{g}/\text{m}^3$   $\text{PM}_{2.5}$ ). When we took individual and area level covariates into account, we found no evidence of clustering in the neighbourhoods (data not shown).

Effect modification by age showed stronger  $\text{PM}_{2.5}$  effects for those aged 60-74 (hazard ratio 1.25, 95% confidence interval 1.03 to 1.51) and for those >75 (1.18, 0.85 to 1.64) than among those aged under 60 (0.91, 0.71 to 1.15), with  $P=0.11$  for effect modification. For all other effect modifiers, the P value of effect modification was far from significant ( $\geq 0.22$ ).

Table 7 shows the results of the threshold analysis for  $\text{PM}_{2.5}$  and  $\text{PM}_{10}$ . When we restricted the analysis to participants from nine cohorts exposed to  $\text{PM}_{2.5}$  concentrations below the current European limit value (<25  $\mu\text{g}/\text{m}^3$ ), there was a 18% increased risk per 5  $\mu\text{g}/\text{m}^3$  increase in  $\text{PM}_{2.5}$  (hazard ratio 1.18, 95% confidence interval 1.01 to 1.39). Similarly, for  $\text{PM}_{10}$  below the current European limit value (40  $\mu\text{g}/\text{m}^3$ ) there was a 12% increased risk per 10  $\mu\text{g}/\text{m}^3$  increase in  $\text{PM}_{10}$  (1.12, 1.00 to 1.27). The effect was present even at exposure levels <15  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{2.5}$  and <20  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{10}$ . This result was not an artefact caused by cohort selection at different thresholds as it persisted when we restricted the analysis to the seven cohorts with data available in each threshold (table 7, right columns).

## Discussion

This European multicentre study found that long term exposure to particulate air pollution is associated with an increased risk

of first coronary events. For both particulate matter with diameter  $<10\ \mu\text{m}$  and  $<2.5\ \mu\text{m}$  ( $\text{PM}_{10}$  and  $\text{PM}_{2.5}$ ), we found associations at levels below the current European limits. The results were substantially robust to adjustment for confounding and model specification.

## Comparison with other studies

Our effect estimates are similar to or higher than those from two cohort studies from the US. For an increase of  $10\ \mu\text{g}/\text{m}^3$  in  $\text{PM}_{2.5}$ , Miller and colleagues found a 21% (hazard ratio 1.21, 95% confidence interval 1.04 to 1.42) increased risk of first coronary heart disease in postmenopausal women, and a (non-significant) 6% higher risk of myocardial infarction.<sup>9</sup> In the US Nurses' Health Study,<sup>11</sup> there was an increased incidence of coronary heart disease for a  $10\ \mu\text{g}/\text{m}^3$  increase in  $\text{PM}_{2.5}$  (hazard ratio 1.11, 0.79 to 1.55) and for a  $10\ \mu\text{g}/\text{m}^3$  increase in coarse particles (1.04, 0.82 to 1.32). In the UK, Atkinson and colleagues found a non-significant 1% (-2% to 5%) higher risk in incident myocardial infarction per  $3\ \mu\text{g}/\text{m}^3$  increase in  $\text{PM}_{10}$ .<sup>14</sup> In the California Teachers Study, Lipsett and colleagues found no evidence of increased myocardial infarction incidence per  $10\ \mu\text{g}/\text{m}^3$  increments in  $\text{PM}_{2.5}$  and  $\text{PM}_{10}$ .<sup>12</sup> The results are also in line with the evidence produced so far by studies on the short term effects of particulate matter on ischaemic heart disease.<sup>35 36</sup>

We found a suggestion of a higher risk associated with particulate exposure in participants aged over 60. Evidence on which age range is the most susceptible is mixed. Miller and colleagues found a higher risk in those aged  $>70$  compared with  $<70$ ,<sup>9</sup> while Gan and colleagues found higher risks in first admissions to hospital for coronary heart disease in both those aged  $<60$  and 60-69 than those aged  $\geq 70$ .<sup>5</sup>

We found an effect of  $\text{PM}_{2.5}$  below 25, 20, and  $15\ \mu\text{g}/\text{m}^3$ , and our results highlight the possible harmful health effects of fine particles well below the current EU annual limit for  $\text{PM}_{2.5}$ . It is clear that generalisability of these findings to the entire European population is not straightforward because the cohorts are not representative of the European population. Besides the differences among the cohorts in age range and underlying risk profiles, however, we detected no heterogeneity in the effect estimates among the cohorts.

There are several possible mechanisms through which exposure to air pollution can affect the cardiovascular system, including systemic inflammation, systemic oxidative stress, thrombosis and coagulation, changes in blood pressure, progression of atherosclerosis, and reduced heart rate variability.<sup>15</sup> Most of this evidence comes from studies of short term exposures and short term responses of cardiovascular function. The pathways implicated provide plausible biological mechanisms based on repeated exacerbations of cardiovascular risk factors potentially leading to long term progression of coronary artery disease.<sup>15 37</sup>

It is unlikely that our findings are driven by an excess in coronary disease mortality because in these and some other cohorts participating in the ESCAPE project we could not find much evidence of an association between exposure to air pollution and mortality for ischaemic heart disease or myocardial infarction.<sup>38</sup>

Our main positive results for coronary events are for an association with particulate matter, and the evidence for other pollutants such as nitrogen oxides could be considered as suggestive based on the small positive estimates. Several predictor variables explained the spatial variation of  $\text{PM}_{10}$  and  $\text{PM}_{2.5}$  in ESCAPE study areas,<sup>30</sup> including not only traffic variables but also population density, industrial sources, urban green, and altitude. Estimated  $\text{PM}_{10}$  at place of residence

therefore represents not only traffic emissions. In contrast, the spatial variation in  $\text{PM}_{2.5}$  absorbance was more exclusively explained by traffic variables. Our findings suggest that sources other than vehicular traffic can have an important role on the risk of acute ischaemic heart disease.

## Strengths and weaknesses

In this first multicentre European study to investigate long term effects of air pollution on incidence of coronary events, we used a standardised approach for assessment of exposure with respect to measurements and modelling.<sup>28-31</sup> Both the measurements and the land use regression models were centrally validated, and data analyses were centrally planned and conducted in the same way in each cohort. The European locations captured in this project provide a wide range of exposures and mixtures that vary from north to south. Although the design of the study did not allow us to benefit from the large variability of pollutant exposure between cohorts, we took advantage of the exposure variation within cohorts. The use of existing cohorts, not designed for air pollution epidemiology, might be seen as a limitation of the ESCAPE project. Indeed, we were able to perform a planned meta-analysis and make the most of existing studies in an efficient way. Therefore, we consider these data the first systematic assessment of the impact of ambient particles on incident coronary artery disease in Europe, highlighting that the burden of disease might be underestimated when estimates of mortality are considered alone.<sup>15 39</sup>

The assessment of exposure was conducted in 2008-11, while the cohorts were enrolled over 15 years, starting in 1992. The use of modelled concentrations based on measurements taken long after the beginning of the study assumes that spatial contrasts were stable over time, an assumption that has been supported from observations in different settings.<sup>40-42</sup> In the ESCAPE project considerable effort was made to take this problem into account; when possible, we back-extrapolated air pollution concentrations.<sup>17</sup> Analyses of the association between back-extrapolated concentrations and incidence of coronary events made no important difference.

We collected individual data on many cardiovascular risk factors to deal with confounding. When we compared the results of the "base model" with a model containing additional cardiovascular risk factors, we obtained similar results. It should be noted that variables used to adjust for area level socioeconomic status were different from cohort to cohort, ranging from the percentage of low income inhabitants in a 5 km grid in KORA to a census block deprivation index in the Italian cohorts. We could not determine whether this might have affected the results, but general adjustment for area level socioeconomic indicators did not have much influence in cohort specific analyses. Data on the most relevant cardiovascular risk factors (smoking, diabetes or hypertension, BMI, physical activity) were available for almost all cohorts, therefore a strong bias in the effect estimates caused by confounding is unlikely. Nevertheless, we cannot rule out residual confounding by, for example, differences in diet. In addition, we observed evidence for effect modification, but even within this large multicentre study, the power to reliably detect effect modification is limited and therefore we selected a limited set of interactions to evaluate a priori.

All cohorts, with the exception of the DCH study, enrolled fewer than 10 000 participants, making the statistical power to detect effects of air pollution in single studies limited. The characteristics of participants were heterogeneous, with two cohorts having younger participants than the others. Younger age could be a reason for the smaller effects found in some of

the cohorts. Finally, in nine out of 11 cohorts outcomes were ascertained from routine administrative databases, which could have led to a less precise effect estimate from misclassification of the outcome.

When we included only the studies with the best exposure assessment models (performance of the land use regression leave-one-out cross validation  $R^2 > 60\%$ ), we found increased effect estimates. This could be because of reduced misclassification of exposure. The cohorts with validation  $R^2 \leq 60\%$ , however, had young participants and (in particular the Italian ones) had higher rates of smoking and a better exposure assessment might be not the only explanation.

## Conclusions

Our study suggests an association between long term exposure to inhalable particulate matter and incidence of coronary events. These associations remained for exposure concentrations below the current European limits. The results of this study, together with other ESCAPE findings, support lowering of European limits for particulate air pollution to adequately protect public health.

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local cohort data; CG contributed to exposure assessment and provided local cohort data; BH, ORN, GP, and PV contributed to the design and provided local cohort data; MK contributed to the exposure assessment and the data analyses; KTE, DH, TL, AR, WS, and KdH contributed to exposure assessment; GW and KW contributed to the statistical script; BB and APe contributed to the design and drafted the manuscript. All authors contributed to critical reading of and comments to the manuscript, interpretation of data and approved the final draft. GC is guarantor.

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Ethical approval: The original cohort studies were approved by appropriate institutional review boards complying with all relevant national, state, and local regulations.

Transparency statement: The lead author affirms that this manuscript is an honest, accurate, and transparent account of the study being reported; that no important aspects of the study have been omitted; and that any discrepancies from the study as planned have been explained.

Data sharing: meta-analytic data and statistical code are available from the corresponding author.

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**What is already known on this topic**

Ambient particulate matter air pollution is estimated to cause 3.1 million deaths worldwide per year, and 22% of DALYs (disability adjusted life years) from ischaemic heart disease

The association between long term exposure to air pollution and incidence of coronary events remains controversial

**What this study adds**

There is an association between outdoor particulate matter and incidence of acute coronary events, even for exposure levels below the current European limits

The burden of disease attributable to outdoor particulate matter might be underestimated if only estimates of mortality are considered

The results of this study support lowering of the EU limits for particulate matter air pollution

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## Tables

**Table 1 | Summary description of 11 European cohorts from five countries contributing data to study long term exposure to ambient air pollution and incidence of acute coronary events**

Variable	FINRISK	SNAC-K	SALT	60 year olds	SDPP	DCH	HNR	KORA	EPIC Turin	SIDRIA	
										Turin	Rome
No of participants	9995	2684	6084	3686	7723	35 693	4433	8301	7230	5137	9200
Person years at risk	105 060	16 256	51 756	39 978	106 995	464 055	34 941	84 595	91 490	56 366	102 894
% of original cohort	93.5	89.4	92.3	91.0	97.6	96.0	99.3	94.0	82.4	95.5	87.1
No of cases	212	200	204	165	181	3293	135	282	157	123	205
Study location	Turku and Helsinki (Finland)	Stockholm (Sweden)	Stockholm (Sweden)	Stockholm (Sweden)	Stockholm (Sweden)	Copenhagen (Denmark)	Ruhr Area (Germany)	Augsburg (Germany)	Turin (Italy)	Turin (Italy)	Rome (Italy)
Years of enrolment	1992, 1997, 2002, 2007	2001-04	1998-2002	1997-99	1992-98	1993-97	2000-03	1994-95, 1999-2001	1993-98	1999	1999



**Table 2 | Individual baseline characteristics from 11 European cohorts from five countries contributing data to study long term exposure to ambient air pollution and incidence of acute coronary events. Values are numbers (percentages) unless stated otherwise**

Variable	FINRISK	SNAC-K	SALT	60 year olds	SDPP	DCH	HNR	KORA	EPIC Turin	SIDRIA	
										Turin	Rome
No of participants	9995 (100)	2684 (100)	6084 (100)	3686 (100)	7723 (100)	35 693 (100)	4433 (100)	8301 (100)	7230 (100)	5137 (100)	9200 (100)
Mean (SD) age (years)	48 (13)	74 (11)	59 (11)	60 (0)	47 (5)	57 (4)	59 (8)	49 (14)	50 (8)	44 (6)	44 (6)
Women	5460 (55)	1751 (65)	3557 (58)	1967 (53)	4721 (61)	19 302 (54)	2309 (52)	4275 (51)	3472 (48)	2677 (52)	4859 (53)
Marital status:											
Single	1601 (16)	397 (15)	835 (14)	170 (5)	1272 (16)*	2352 (7)	256 (6)	879 (11)	438 (6)	121 (2)	0 (0)
Married/living with partner	7009 (70)	1263 (47)	4103 (67)	2628 (71)	6451 (84)	24 673 (69)	3319 (75)	6292 (76)	6183 (86)	4897 (95)	9200 (100)
Divorced/separated	1051 (11)	362 (13)	678 (11)	630 (17)	—	6599 (18)	444 (10)	620 (7)	375 (5)	37 (1)	0 (0)
Widowed	334 (3)	662 (25)	468 (8)	258 (7)	—	2069 (6)	414 (9)	510 (6)	234 (3)	82 (2)	0 (0)
Education:											
Primary school or less	3029 (30)	705 (26)	1332 (22)	1030 (28)	1996 (26)	10 589 (30)	499 (11)	1024 (12)	3168 (44)	898 (17)†	4130 (45)
Secondary school	5217 (52)	1128 (42)	2606 (43)	1631 (44)	3453 (45)	16 943 (47)	2466 (56)	6200 (65)	3081 (43)	3668 (71)	3690 (40)
University degree and more	1749 (17)	851 (32)	2146 (35)	1025 (28)	2274 (29)	8161 (23)	1468 (33)	1077 (13)	981 (14)	571 (11)	1380 (15)
Occupational status:											
Employed/self employed	7092 (71)	2016 (75)	—	1881 (51)	7088 (92)	28 600 (80)	1859 (42)	4908 (59)	—	3727 (73)	6500 (71)
Unemployed	610 (6)	668 (25)‡	—	374 (10)	635 (8)‡	7093 (20)‡	630 (14)	272 (3)	—	349 (7)	393 (4)
Homemaker or housewife	351 (4)	—	—	290 (8)	—	—	1664 (38)	1188 (14)	—	1061 (21)	2307 (25)
Retired	1942 (19)	—	—	1141 (31)	—	—	280 (6)	1933 (23)	—	0 (0)	0 (0)
Smoking status:											
Current smoker	2603 (26)	394 (15)	1224 (20)	751 (20)	2022 (26)	12 793 (36)	1040 (23)	2162 (26)	1768 (24)	2117 (41)	3884 (42)
Former smoker	2808 (28)	927 (34)	2678 (44)	1429 (39)	2815 (36)	9944 (28)	1482 (33)	2517 (30)	2368 (33)	1085 (21)	2147 (23)
Never smoker	4584 (46)	1363 (51)	2182 (36)	1506 (41)	2886 (37)	12 956 (36)	1911 (43)	3622 (44)	3094 (43)	1935 (38)	3169 (34)
Mean (SD) years of smoking among ever smokers	15 (12)	30 (17)	—	26 (13)	20 (10)	29 (10)	36 (9)§	21 (13)	23 (10)	18 (8)	18 (7)
Mean (SD) No of cigarettes/day among current smokers	15 (9)	11 (8)	13 (8)	13 (7)	14 (7)	17 (10)	17 (12)	15 (11)	14 (9)	15 (9)	15 (9)

\*All except married/living with partner.

†&lt;6 years of schooling.

‡All except employed.

§Only among current smokers.

**Table 3| Air pollution exposure at residential address of individuals from 11 European cohorts from five countries contributing data to study long term exposure to ambient air pollution and incidence of acute coronary events. Figures are means and 5-95th centile ranges, correlation coefficients between PM<sub>2.5</sub> and other pollutants, and R<sup>2</sup> of models and their cross validation**

Variable	FINRISK	SNAC-K	SALT	60 year olds	SDPP	DCH	HNR	KORA	EPIC Turin	SIDRIA	
										Turin	Rome
<b>Mean (5th-95th centile) annual air pollution (µg/m<sup>3</sup>) at residence address</b>											
PM <sub>10</sub>	14 (10-20)	16 (6-29)	15 (7-21)	15 (7-21)	14 (6-17)	17 (14-20)	28 (25-32)	20 (16-24)	46 (39-52)	48 (41-54)	36 (31-47)
Coarse PM	7 (4-11)	8 (1-19)	7 (2-12)	7 (1-12)	6 (1-9)	6 (4-7)	10 (7-12)	6 (5-8)	16 (12-20)	17 (13-20)	17 (12-24)
PM <sub>2.5</sub>	8 (6-9)	8 (6-10)	7 (5-9)	7 (5-9)	7 (5-8)	11 (10-13)	18 (17-20)	14 (13-15)	30 (27-33)	31 (29-34)	19 (17-23)
Absorbance PM <sub>2.5</sub> ·10 <sup>5</sup> /m	0.9 (0.5-1.2)	0.8 (0.5-1.2)	0.6 (0.4-0.9)	0.6 (0.4-0.9)	0.5 (0.4-0.7)	1.2 (0.8-1.5)	1.6 (1.2-2.2)	1.7 (1.5-2.0)	3.1 (2.3-3.6)	3.2 (2.6-3.8)	2.7 (2.2-4.0)
NO <sub>2</sub>	15 (9-24)	17 (9-25)	11 (7-20)	11 (6-20)	8 (6-11)	16 (8-30)	30 (23-39)	19 (14-26)	53 (34-68)	60 (42-77)	39 (26-56)
NO <sub>x</sub>	24 (14-41)	33 (15-58)	19 (12-40)	19 (12-39)	14 (12-20)	27 (7-66)	51 (33-72)	33 (24-47)	96 (62-132)	107 (79-162)	82 (39-122)
Daily No of vehicles/day on nearest road	1670 (50-9011)	3726 (500-21828)	1454 (500-6000)	1455 (500-6300)	864 (500-2575)	2994 (200-16145)	-	1636 (500-8367)	3907 (0-23951)	4290 (0-24379)	2966 (500-15312)
Daily traffic load on major roads in 100 m buffer (thousand), (vehicles×meters/day)	633 (0-3711)	2307 (0-6572)	578 (0-3437)	521 (0-3048)	109 (0-986)	1274 (51-4719)	1017 (0-4302)	444 (0-2805)	466 (0-2340)	804 (0-4197)	1417 (0-6947)
<b>Pearson correlations coefficients between PM<sub>2.5</sub> and:</b>											
PM <sub>10</sub>	0.67	0.70	0.49	0.50	0.31	0.74	0.90	0.42	0.62	0.56	0.92
Coarse PM	0.10	0.71	0.50	0.50	0.32	0.60	0.51	0.38	0.51	0.32	0.90
Absorbance PM <sub>2.5</sub>	0.98	0.98	0.84	0.84	0.90	0.49	0.76	0.50	0.77	0.73	0.78
NO <sub>2</sub>	0.41	0.82	0.60	0.61	0.61	0.57	0.63	0.45	0.72	0.67	0.69
<b>R<sup>2</sup> of land use regression models and their leave one out cross validation</b>											
PM <sub>10</sub>	0.67-0.42	0.82-0.77	0.82-0.77	0.82-0.77	0.82-0.77	0.75-0.64	0.69-0.63	0.83-0.75	0.78-0.69	0.78-0.69	0.72-0.59
Coarse PM	0.61-0.33	0.72-0.65	0.72-0.65	0.72-0.65	0.72-0.65	0.71-0.54	0.66-0.57	0.81-0.79	0.65-0.58	0.65-0.58	0.70-0.57
PM <sub>2.5</sub>	0.67-0.53	0.87-0.78	0.87-0.78	0.87-0.78	0.87-0.78	0.62-0.55	0.88-0.79	0.78-0.62	0.71-0.59	0.71-0.59	0.71-0.60
Absorbance PM <sub>2.5</sub>	0.65-0.47	0.89-0.85	0.89-0.85	0.89-0.85	0.89-0.85	0.92-0.86	0.97-0.95	0.91-0.82	0.88-0.81	0.88-0.81	0.84-0.70
NO <sub>2</sub>	0.83-0.75	0.82-0.78	0.82-0.78	0.82-0.78	0.82-0.78	0.88-0.83	0.89-0.84	0.86-0.67	0.78-0.70	0.78-0.70	0.87-0.76
NO <sub>x</sub>	0.85-0.74	0.83-0.79	0.83-0.79	0.83-0.79	0.83-0.79	0.83-0.73	0.88-0.81	0.88-0.76	0.78-0.72	0.78-0.72	0.80-0.79

PM<sub>2.5</sub>=particulate matter <2.5 µm; PM<sub>10</sub>=particulate matter <10 µm; NO<sub>2</sub>=nitrogen dioxide; NO<sub>x</sub>=nitrogen oxides.

**Table 4** Additional baseline characteristics of study populations in analysis of long term exposure to ambient air pollution and incidence of acute coronary events. Values are numbers (percentages) unless stated otherwise

	FINRISK	SNAC-K	SALT	60 year olds	SDPP	DCH	HNR	KORA	EPIC Turin	SIDRIA	
										Turin	Rome
Area level socioeconomic indicators <sup>2</sup>											
1 (lowest)	—	1429 (53)	1692 (28)	898 (24)	3024 (39)	—	—	—	963 (13)	1114 (22)	2332 (25)
2	—	1252 (47)	2643 (43)	924 (25)	1164 (15)	—	—	—	1404 (19)	1114 (22)	1802 (20)
3	—	3 (0)	242 (4)	938 (25)	2016 (26)	—	—	—	1342 (19)	944 (18)	1675 (18)
4	—	—	1507 (25)	926 (25)	1519 (20)	—	—	—	1662 (23)	1070 (21)	1684 (18)
5 (highest)	—	—	—	—	—	—	—	—	1859 (26)	895 (17)	1707 (19)
Mean (SD)	22 946 (5458)					1.9 (0.4)	12.6 (3.1)	28.1 (18.4)			
Mean (SD)	26.3 (4.6)	25.6 (4.1)	28.5 (4.1)	26.8 (4.2)	25.7 (4.0)	26.0 (4.1)	27.8 (4.6)	27.1 (4.6)	25.3 (3.8)	—	—
BMI											
BMI in classes:											
<25	4329 (43)	1209 (48)	1055 (18)	1349 (37)	3741 (49)	15 998 (45)	1201 (27)	2824 (34)	3691 (51)	—	—
25-29	3852 (39)	1004 (40)	3131 (52)	1633 (44)	2995 (39)	14 642 (41)	2025 (46)	3551 (43)	2779 (38)	—	—
>29	1811 (18)	302 (12)	1805 (30)	704 (19)	964 (13)	5026 (14)	1189 (27)	1852 (23)	761 (11)	—	—
Physical activity (hours/week):											
<1	1808 (18)	464 (20)	1575 (26)	2510 (69)	846 (11)	15 984 (45)	2263 (51)	3107 (37)	1660 (23)	—	—
~1	3322 (33)	686 (30)	3757 (62)	858 (24)	6273 (81)	19 709 (55)	490 (11)	3515 (42)	2032 (28)	—	—
>2	4839 (49)	1115 (49)	716 (12)	279 (8)	597 (8)	—	1667 (38)	1666 (20)	3538 (49)	—	—
Alcohol consumption (drinks/week):											
Never	1320 (13)	519 (19)	—	168 (5)	594 (8)	776 (2)	1015 (23)	3867 (47)	457 (6)	—	—
1-3	4566 (46)	634 (24)	—	771 (21)	2876 (38)	34 142 (98)†	1190 (27)	1212 (15)	201 (3)	—	—
3-6	3139 (32)	1236 (46)	—	1622 (44)	3873 (51)	—	2127 (49)	549 (7)	347 (5)	—	—
>6	824 (8)	288 (11)	—	1123 (30)	280 (4)	—	—	2662 (32)	6225 (86)	—	—
Diabetes mellitus:											
No	9546 (96)	2464 (92)	5833 (96)	3533 (96)	7597 (98)	34 943 (98)	3877 (87)	7940 (96)	7118 (98)	5111 (99)	9179 (100)
Yes	426 (4)	220 (8)	251 (4)	153 (4)	126 (2)	703 (2)	556 (13)	360 (4)	110 (2)	26 (1)	21 (0)
Hypertension:											
No	5877 (59)	902 (34)	4735 (78)	1770 (48)	5801 (76)	30 033 (84)	2002 (45)	5098 (62)	3629 (53)	5082 (99)	9124 (99)
Yes	4042 (41)	1762 (66)	1347 (22)	1915 (52)	1837 (24)	5625 (16)	2423 (55)	3190 (38)	3255 (47)	55 (1)	76 (1)
Mean (SD)	211 (78)	—	—	231 (41)	—	—	231 (39)	229 (44)	—	—	—
serum cholesterol											
Noise exposure at baseline address (dB):											
<45			123 (5)	84 (6)	—	126 (0)	721 (17)	466 (6)	44 (1)	62 (1)	—
45-49	1912 (23)‡	319 (12)	200 (8)	110 (7)	—	1838 (5)	989 (23)	1345 (16)	115 (2)	121 (2)	—
50-54	2040 (24)	799 (30)	524 (21)	292 (20)	—	9634 (27)	853 (20)	2946 (36)	87 (2)	103 (2)	—
55-59	1721 (21)	692 (26)	550 (23)	304 (21)	—	9560 (27)	565 (13)	1873 (23)	253 (5)	186 (4)	—
60-64	1181 (14)	627 (24)	435 (18)	291 (20)	—	7625 (21)	482 (11)	987 (12)	2419 (43)	2237 (44)	—
65-69	784 (9)	230 (9)	332 (14)	225 (15)	—	4600 (13)	462 (11)	478 (6)	1254 (22)	1133 (22)	—
70-74	714 (9)‡	0	237 (10)	143 (10)	—	2010 (6)	194 (5)	199 (2)	1369 (24)	1126 (22)	—
>74	—	—	43 (2)	30 (2)	—	294 (1)	40 (1)	3 (0)	136 (2)	156 (3)	—
Living in areas of low urbanisation:											
No	9086 (91)	2684 (100)	2763 (45)	1450 (39)	0 (0)	14 155 (40)	4433 (100)	3557 (43)	7230 (100)	5137 (100)	9200 (100)
Yes	909 (9)	0 (0)	3321 (55)	2236 (61)	7723 (100)	21 538 (60)	0 (0)	4744 (57)	0 (0)	0 (0)	0 (0)

\*FINRISK: median income rate in 3 km area; SNAC-K: mean income in thirds at small neighbourhood level (Small Area for Market Statistics); SALT and SDPP: mean income in four categories at municipality level; 60 years: mean income in quarters at small neighbourhood level (Small Area for Market Statistics); DCH: mean income at municipality level (16 units with median population of 1500 inhabitants), per/100 000; HNR: unemployment rate at neighbourhood level; KORA: percentage of low income in 5 km grid; EPIC-Turin, SIDRIA-Turin, and SIDRIA-Rome: deprivation index, census block level (average 500 inhabitants).

**Table 4 (continued)**

	FINRISK	SNAC-K	SALT	60 year olds	SDPP	DCH	HNR	KORA	EPIC Turin	SIDRIA	
										Turin	Rome

†Any category of drinking.

‡First and last two classes are collapsed.



**Table 5 | Association between exposure to pollutants and incidence of coronary events. Results expressed per fixed increments**

Exposure	Fixed increment	Cohorts	Participants	Pooled hazard ratios (95% CI)		
				Model 1*	Model 2†	Model 3‡
PM <sub>10</sub> (µg/m <sup>3</sup> )	10	11	100 166	1.15 (1.04 to 1.28)	1.12 (1.01 to 1.24)	1.12 (1.01 to 1.25)
Coarse PM (µg/m <sup>3</sup> )	5	11	100 166	1.08 (1.00 to 1.17)	1.06 (0.98 to 1.14)	1.06 (0.98 to 1.15)
PM <sub>2.5</sub> (µg/m <sup>3</sup> )	5	11	100 166	1.22 (1.04 to 1.44)	1.15 (1.00 to 1.32)	1.13 (0.98 to 1.30)
Absorbance PM <sub>2.5</sub> (10 <sup>-5</sup> /m)	1	11	100 166	1.18 (1.05 to 1.32)	1.10 (0.98 to 1.24)	1.10 (0.98 to 1.24)
NO <sub>2</sub> (µg/m <sup>3</sup> )	10	11	100 166	1.04 (0.96 to 1.12)	1.03 (0.96 to 1.11)	1.03 (0.97 to 1.08)
NO <sub>x</sub> (µg/m <sup>3</sup> )	20	11	100 166	1.03 (1.00 to 1.07)	1.01 (0.98 to 1.05)	1.01 (0.98 to 1.05)
Traffic intensity on nearest road adjusted for background NO <sub>2</sub> (vehicles/day)§	5000	10	95 733	1.01 (0.98 to 1.04)	1.01 (0.98 to 1.04)	1.01 (0.98 to 1.04)
Traffic load on major roads in 100 m buffer adjusted for background NO <sub>2</sub> (vehicles×m/day)	4 000 000	11	100 166	1.02 (0.96 to 1.08)	1.00 (0.95 to 1.06)	1.00 (0.95 to 1.06)

\*Adjusted for age (time variable), year of enrolment and sex

†As for model 1 plus adjusted for marital status, education, occupation, smoking status, smoking duration, and smoking intensity.

‡As for model 2 plus adjusted for socioeconomic area level variables.

§All cohorts except HNR study.<sup>24</sup>

**Table 6 | Incidence of coronary events, results from sensitivity analyses. Figures are pooled hazard ratios (and 95% CI) for 10 µg/m<sup>3</sup> PM<sub>10</sub> and 5 µg/m<sup>3</sup> PM<sub>2.5</sub>, I<sup>2</sup> statistics, and P value for heterogeneity**

Model	No of cohorts	No of participants*	Hazard ratio† (95% CI)	
			10 µg/m <sup>3</sup> PM <sub>10</sub>	5 µg/m <sup>3</sup> PM <sub>2.5</sub>
Base model (model 3)	11	100 166	1.12 (1.01 to 1.25)	1.13 (0.98 to 1.30)
<b>Role of cardiovascular risk factors</b>				
Intermediate variables: diabetes and hypertension:				
+ diabetes and hypertension	11	99 526	1.11 (1.00 to 1.24)	1.11 (0.96 to 1.28)
Physical activity, alcohol and BMI‡:				
Base model, on subset of participants with additional information	8	77 584	1.17 (1.04 to 1.33)	1.22 (1.03 to 1.44)
+ additional information	8	77 584	1.17 (1.04 to 1.33)	1.19 (1.01 to 1.42)
All cardiovascular risk factors (diabetes, hypertension, physical activity, alcohol, BMI, cholesterol)§:				
Base model, on subset of participants with additional information	4	25 625	1.19 (0.92 to 1.53)	1.28 (0.92 to 1.79)
+ additional information	4	25 625	1.20 (0.93 to 1.56)	1.29 (0.92 to 1.81)
<b>Urban/suburban-rural residence location:</b>				
Base model, on subset of participants with additional information	11	100 166	1.12 (1.01 to 1.25)	1.13 (0.98 to 1.39)
+ additional information	11	100 166	1.13 (1.02 to 1.26)	1.14 (0.99 to 1.31)
<b>Noise¶:</b>				
Base model, on subset of participants with noise information	9	73 840	1.14 (1.01 to 1.30)	1.16 (0.98 to 1.37)
+ noise variable	9	73 840	1.14 (1.01 to 1.31)	1.13 (0.95 to 1.36)
<b>Change of address**:</b>				
Base model, on subset of cohorts with change of address information	10	92 936	1.11 (0.99 to 1.23)	1.14 (0.98 to 1.33)
Participants who didn't change address during follow-up	10	63 121	1.16 (1.01 to 1.32)	1.18 (0.98 to 1.42)
<b>Proportionality hazards (PH) assumption:</b>				
Variables that do not meet PH put as strata	11	100 166	1.10 (0.99 to 1.23)	1.13 (0.98 to 1.30)
<b>Exclusion of DCH Study</b>				
Data after exclusion	10	64 473	1.12 (0.99 to 1.27)	1.09 (0.92 to 1.30)
<b>Performance of land use regression model††:</b>				
Leave one out cross validation R <sup>2</sup> >60%	—	—	1.18 (1.05 to 1.33)	1.35 (1.04 to 1.74)
Leave one out cross validation R <sup>2</sup> ≤60%	—	—	0.88 (0.68 to 1.12)	1.05 (0.89 to 1.25)

\*Figures depend on number of cohorts included and on missing values of included variables

†Adjusted for age (time variable), year of enrolment, sex, marital status, education, occupation, smoking status, smoking duration, smoking intensity, and socioeconomic area level indicators.

‡Excludes SALT, SIDRIA-Turin, and SIDRIA-Rome.

§Includes FINRISK, 60 year olds, HNR, and KORA.

¶All cohorts except SDPP and SIDRIA-Rome.

\*\*All cohorts except EPIC-Turin.

††For PM<sub>10</sub>: nine cohorts, 80 971 participants (excludes FINRISK and SIDRIA-Rome). For PM<sub>2.5</sub>: six cohorts, 42 906 participants (excludes FINRISK, DCH, EPIC-Turin, SIDRIA-Turin, and SIDRIA-Rome).

**Table 7 | Results from random-effects meta-analyses for adjusted association between incidence of coronary events and exposure to PM10 and PM2.5 below various threshold values. Figures are pooled hazard ratios (95% CI) for 10 µg/m<sup>3</sup> PM10 and 5 µg/m<sup>3</sup> PM2.5, I<sup>2</sup> statistic, and P value for heterogeneity**

Exposure and threshold	Cohorts with exposure data available for respective threshold					Cohorts with exposure data available for all thresholds				
	No of cohorts	No of participants	HR* (95% CI)	I <sup>2</sup>	P value	No of cohorts	No of participants	HR* (95% CI)	I <sup>2</sup>	P value
<b>PM<sub>10</sub> (µg/m<sup>3</sup>):</b>										
Full range of exposure	—	—	—	—	—	7†	74 166	1.15 (1.02 to 1.30)	0	0.92
<20	7†	64 363	1.20 (1.01 to 1.41)	0	0.91	7†	64 363	1.20 (1.01 to 1.41)	0	0.91
<30	8‡	77 997	1.12 (0.98 to 1.27)	0	0.94	7†	74 019	1.12 (0.98 to 1.27)	0	0.91
<40	9§	86 222	1.12 (1.00 to 1.27)	0	0.48	7†	74 166	1.15 (1.02 to 1.30)	0	0.92
<b>PM<sub>2.5</sub> (µg/m<sup>3</sup>):</b>										
Full range of exposure	—	—	—	—	—	7†	74 166	1.23 (1.04 to 1.46)	0	0.57
<15	7†	73 420	1.19 (1.00 to 1.42)	0	0.74	7†	73 420	1.19 (1.00 to 1.42)	0	0.74
<20	9§	85 216	1.17 (0.91 to 1.50)	30	0.18	7†	74 166	1.23 (1.04 to 1.46)	0	0.57
<25	9§	87 532	1.18 (1.01 to 1.39)	0	0.56	7†	74 166	1.23 (1.04 to 1.46)	0	0.57

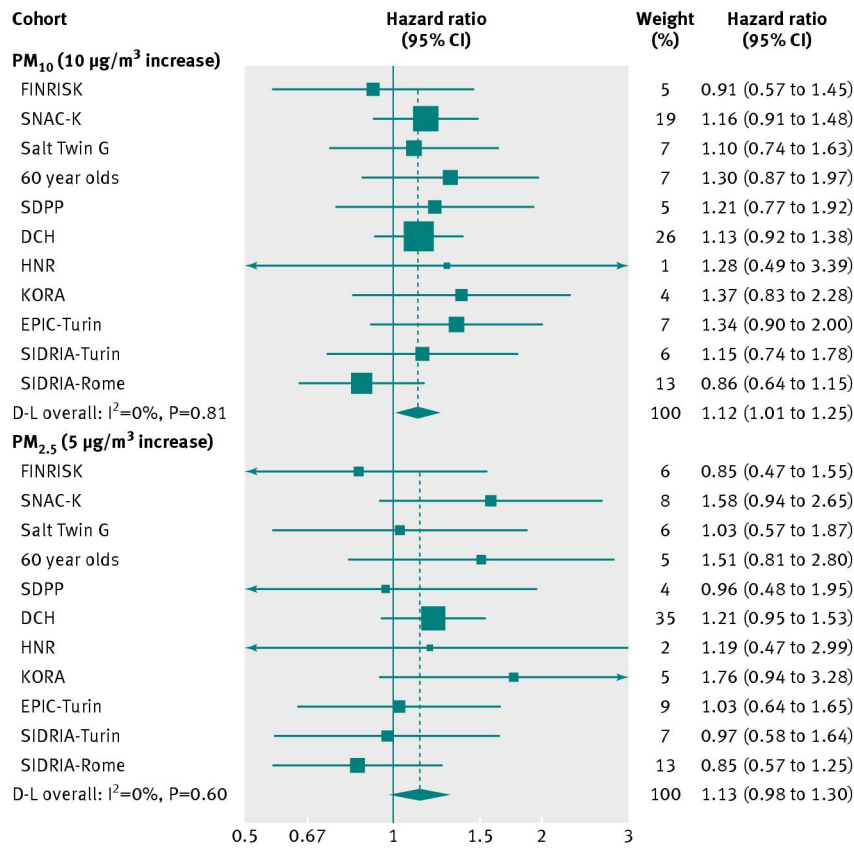
\*Adjusted for age (time variable), year of enrolment, sex, marital status, education, occupation, smoking status, smoking duration, smoking intensity, and socioeconomic area level indicators

†All cohorts except Heinz Nixdorf Recall Study, EPIC-Turin, SIDRIA-Turin and SIDRIA-Rome.

‡All cohorts except EPIC-Turin, SIDRIA-Turin and SIDRIA-Rome.

§All cohorts except EPIC-Turin and SIDRIA-Turin.

# Figure



Hazard ratios of incident coronary events per 10 µg/m<sup>3</sup> PM<sub>10</sub> and 5 µg/m<sup>3</sup> PM<sub>2.5</sub>. Cohort specific and meta-analytic results